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# Quantitative structure-toxicity relationship study of some polychlorinated aromatic compounds using molecular descriptors 

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#### Abstract

Polychlorinated aromatic compounds represent a large group of industrial and byproduct compounds which are resistant to chemical and biological degradation and highly toxic. QSAR analysis was performed on 74 molecules of three classes of polychlorinated aromatic compounds (polychlorinated dibenzo-p-dioxin (PCDDs), polychlorinated dibenzofuran (PCDF) and polychlorinated biphenyl (PCB)). A large number of about 1700 molecular descriptors was obtained from DFT (B3LYP/6-311+G*) level of calculation for each molecule and used in Genetic function algorithm (GFA) approach to generate 5 models, out of which the one with the highest statistical significance (Model-1: $R^{2}=0.9673, R_{\text {adjusted }}^{2}=0.9592, R_{c v}^{2}=0.9402, R_{\text {pred }}^{2}=0.7209, F$-test $=118.48, L O F=0.4377$ ) was selected as the best. From the model generated, it seems to be very clear that polarizability, SP-7, ETA_Epsilon_5, GRAVH_3, and MOMI-R contribute positively to the toxicity of these compounds while MaxHBint5, ETA_dApha_B, ETA_Epsinlon-2, n5Ring and GRAV_2 contribute negatively. This validated model brings important insight to aid the prediction and identification of other toxic polychlorinated aromatic compounds.


Keywords: QSAR, Genetic Function Algorithm, Molecular descriptors, Polychlorinated aromatic compounds, toxicit of polychlorinated compounds.

## INTRODUCTION

In recent years, due to the increasing impact the pharma and food chemistry in special have on the human and environment life, the scientific and economical interest forced the international communities (OECD- Organization of economic cooperation and development, EUC-European Commissions, just to name $a$ few) to adopt memorandums regulating the design and use of chemicals towards lower toxicity and higher biodegrability [1]. Polychlorinated dibenzofurans (PCDFs), polyhalogenateddibenzo-p-dioxins (PHDDs) and polychlorinated biphenyls (PCBs) are chemicals of concern because of their elevated concentrations in adipose and hepatic tissues and their persistence in an individual for extended lengths of time. With heavier congeners, it may stay with an individual for decades because they are resistant to metabolic, thermal and environmental breakdown. Polychlorinated aromatic compounds are not commercially produced but are formed as trace amounts of undesired impurities in the manufacture of other chemicals [2-3].

The chemical/industrial sources of these chemicals include the manufacture of chlorinated compounds such as phenoxy herbicides, chlorinated benzenes, chlorinated aliphatic compounds, chlorinated catalysts and halogenated diphenyl ethers, PCBs, the pulp and paper industry, and dry cleaning distillation residues [4]. They could also be produced when organic compounds containing chlorine are burned and a series of chemical reactions take place under specific conditions [5]. These combustion sources include incinerators for municipal solid waste and hazardous waste, steelworks, metal refinery factories, power stations, coal and oil industries, sintering plants, cement, lime, glass and brick production, and recycling plants [6,7].The use and disposal of these compounds can cause the release of dioxins into the environment.

Polychlorinated aromatic compounds are considered as persistent and widespread environmental contaminants with high hydrophobicity, which can cause a great diversity of biological effects including hepatotoxicity, endocrine effects, immunotoxicity, body weight loss, teratogenicity, carcinogenicity and the induction of diverse enzymes such a aryl hydrocarbon hydroxylase (AHH) and 7-ethoxyresorufinOdeethylase (EROD) in various organisms [8,9].

Food is the major source for human exposure to PCBs and dioxins, especially fatty foods: dairy products (butter, cheese, fatty milk), meat, egg, and fish. Food of animal origin accounts for $95 \%$ of total exposure. The current average body burden of dioxins is about $5-50 \mathrm{ng} / \mathrm{kg}$ (as WHO TEq in fat; $\mathrm{pg} / \mathrm{g}=\mathrm{ng} / \mathrm{kg}$ ) or $100-1000 \mathrm{ng}$ (WHO-TEq) per person which is close to the lowest concentrations possibly causing health effects. Some subgroups within the society (e.g., nursing babies and people consuming plenty of fish) may be exposed to higher than average amounts of these compounds and are thus at greater risk. Dioxin concentrations have been screened in five WHO international studies, and in Central Europe the concentrations have decreased in breast milk from about $40 \mathrm{ng} / \mathrm{kg}$ (as TEq in milk fat) in 1987 to below $10 \mathrm{ng} / \mathrm{kg}$ in 2006. PCBs have decreased at about the same rate. The decrease in environmental concentrations is due to cessation of PCB use and improved incineration technology [10]. Due to the problems of assessing the fate and toxicity of large number of chemicals, alternative method has been sought to classical in vivo animal texting. In the area of computer - aided toxicity prediction, quantitative structure activity relationship (QSAR) have been seen as an attractive method for toxicity and fate assessment [11].The study of the quantitative relationship between toxicity/activity and molecular structure (QSTR/QSAR) is an important area of research in computational chemistry and has been widely used in the prediction of toxicity and other biological activities of organic compounds [12, 13].

In this study, genetic function approximation (GFA) which is a statistical modeling algorithm that builds functional models of experimental data. Since its inception, several applications of this algorithm in the area of quantitative structure-activity relationship modeling have been reported [14]. The genetic function approximation (G FA) algorithm is a genetic algorithm (GA) derived from the previously reported G/SPLINES algorithm and has been recently applied to the generation of QSAR models [15-17]. The main purpose of this work is to find out how accurate QSAR analysis (using Material studio 7.0 software and the statistical tool Genetic functional algorithm) predicted the toxicity of polychlorinated aromatic compounds, and also to find out the descriptors responsible for producing such toxicity other than the once reported by $[18,19]$.

## MATERIALS AND METHODS

## QSAR METHODOLOGY

## Chemical data and biological activity

A data set of 74 molecules ( 25 PCDDs, 34 PCDFs and 15 PCBs ) has been taken from the literature [20-22]. The toxicities of the compounds expressed in $\mathrm{EC}_{50}$ have been converted to $\log \mathrm{EC}_{50}$. The structures of the compounds were drawn using Chemsketch software. The general structural formulae of the three series are shown in Fig.2. IUPAC names and toxicity data of all the compounds are listed in Table-1.


Polychlorinated dibenzo-P-dioxins (PCDDs)


Polychlorinated biphenyls (PCBs)


Polychlorinated dibenzofurans (PCDFs)
Fig.2: General structural formulae for the compounds (PCDDs, PCDFs and PCBs)
Table 1: Experimental Biological Activities in $\mathrm{pEC}_{50}$ of the compounds, (PCDDs, PCDFs and PCBs)

| S/N | IUPAC NAME | pEC 50 |
| :---: | :---: | :---: |
| 1 | 2,3,7,8 Tetrachlorodibenzo-p-dioxin | 8.00 |
| 2 | 1,2,3,7,8 pentachlorodibenzo-p-dioxin | 7.10 |
| 3 | 2,3,6,7 Tetrachlorodibenzo-p-dioxin | 6.80 |
| 4 | 2,3,6 Trichlorodibenzo-p-dioxin | 6.66 |
| 5 | 1,2,3,4,7,8Hexachlorodibenzo-p-dioxin | 6.55 |
| 6 | 1,3,7,8-Tetrachlorodibenzo-p-dioxin | 6.10 |
| 7 | 1,2,4,7,8-pentachlorodibenzo--dioxin | 5.96 |
| 8 | 1,2,3,4-Tetrachlorodibenzo-p-dioxin | 5.89 |
| 9 | 2,3,7-Trichlorodibenzo-p-dioxin | 7.15 |
| 10 | 1,2,3,4,7-pentachlorodibenzo-p-dioxin | 5.19 |
| 11 | 1,2,4-Trichlorodibenzo-p-dioxin | 4.89 |
| 12 | 2,8-dichlorodibenzo-p-dioxine | 5.49 |
| 13 | 1,2,3,4,6,7,8,9-Octachlorodibenzo-o-dioxin | 5.00 |
| 14 | 1-chlorodibenzo-p-dioxin | 4.00 |
| 15 | 2,3,7,8-Tetrabromodibenzo-p-dioxin | 8.82 |
| 16 | 2,3-Dibromo 7,8-chlorodibenzo-p-dioxin | 8.83 |
| 17 | 2,8- Dibromo -3,7-dichlorodibenzo-p-dioxin | 9.35 |
| 18 | 2,Bromo-3,7,8-trichlorodibenzo-p-dioxin | 7.94 |
| 19 | 1,3,7,9-Tetrabromodibenzo-p-dioxin | 7.03 |
| 20 | 1,3,7,8-Tetrabromodibenzo-p-dioxin | 8.70 |
| 21 | 1,2,4,7,8-Pentabromodibenzo-p-dioxin | 7.77 |
| 22 | 1,2,3,7,8-Pentabromodibenzo-p-dioxin | 8.18 |
| 23 | 2,3,7-Tribromodibenzo-p-dioxin | 8.93 |
| 24 | 2,7-Dibromodibenzo-p-dioxin | 7.81 |
| 25 | 2-Bromodibenzo-p-dioxin | 6.53 |
| 26 | 1-chlorodibenzofuran | 4.53 |
| 27 | 2-chlorodibenzofuran | 3.55 |
| 28 | 3-chlorodibenzofuran | 4.38 |
| 29 | 4-chlorodibenzofuran | 3.00 |
| 30 | 2,3-Dichlorodibenzofuran | 5.36 |
| 31 | 2,6-dichlorodibenzofuran | 3.61 |
| 32 | 2,8-Dichlorodibenzofuran | 5.05 |
| 33 | 1,3,6-Trichlorodibenzofuran | 5.36 |
| 34 | 1,3,8-Trichlorodibenzofuran | 4.07 |
| 35 | 2,3,4-Trichlorodibenzofuran | 4.72 |
| 36 | 2,3,7-Trichlorodibenzofuran | 7.10 |
| 37 | 2,3,8-Trichlorodibenzofuran | 6.00 |
| 38 | 2,6,7 Trichlorodibenzofuran | 6.35 |
| 39 | 2,3,4,6 Tetrachlorodibenzofuran | 6.46 |
| 40 | 2,3,4,8-Tetrachlorodibenzofuran | 6.70 |
| 41 | 2,3,7,8-Tetrachlorodibenzofuran | 7.39 |
| 42 | 1,2,4,8-Tetrachlorodibenzofuran | 5.00 |
| 43 | 1,2,4,7,9-Pentachlorodibenzofuran | 4.70 |
| 44 | 1,2,3,7,8-Pentachlorodibenzofuran | 7.13 |


| 45 | 1,2,4,7,8-Pentachlorodibenzofuran | 5.89 |
| :---: | :---: | :---: |
| 46 | 2,3,4,7,8-Pentachlorodibenzofuran | 7.82 |
| 47 | 1,2,3,4,7,8-Hexachlorobenzofuran | 6.64 |
| 48 | 1,2,3,6,7,8-Hexachlorobenzofuran | 6.57 |
| 49 | 2,3,4,67,8-Hexachlorobenzofuran | 7.33 |
| 50 | 1,2,4,6,7,9-Hexachlorodibenzofaran | 5.08 |
| 51 | 2,3,6,8-Tetrachlorodibenzofuran | 6.66 |
| 52 | 1,2,3,6-Tetrachlorodibenzofuran | 6.46 |
| 53 | 1,2,3,7-Tetrachlorodibenzofuran | 6.96 |
| 54 | 1,3,4,7,8-Pentachlorodibenzofuran | 6.70 |
| 55 | 2,3,4,7,9-Pentachlorodibenzofuran | 6.70 |
| 56 | 1,2,3,7,9-Pentachlorodibenzofuran | 6.40 |
| 57 | Dibenzofuran | 3.00 |
| 58 | 2,3,4,7-Tetrachlorobiphenyl | 7.60 |
| 59 | 1,2,4,6,8-Pentachlorobiphenyl | 5.51 |
| 60 | 2,3,4,4'-Tetrachlorobiphenyl | 4.94 |
| 61 | 3,3',4, ''-Tetrachlorobiphenyl | 6.15 |
| 62 | 3,4,4, 5-Tetrachlorobiphenyl | 4.55 |
| 63 | 2',3,4,4',5-Pentachlorobiphenyl | 4.85 |
| 64 | 2,3,3',4,4'-Pentachlorobiphenyl | 5.37 |
| 65 | 2,3',4,4',5-Pentachlorobiphenyl | 5.04 |
| 66 | 2,3,4,4',5-Pentachlorobiphenyl | 5.39 |
| 67 | 3,3'4,4'5-Pentachlorobiphenyl | 6.92 |
| 68 | 2,2'4,4',5,5'-Hexachlorobiphenyl | 4.26 |
| 69 | 2,3,3',4,4',5-Hexachlorobiphenyl | 5.15 |
| 70 | 2,3', 4, 4',5,5'-Hexachlorobiphenyl | 4.80 |
| 71 | 2,3,3',4,4', ${ }^{\prime}$ '-Hexachlorobiphenyl | 5.30 |
| 72 | 2,2', $4,4^{\prime}$-Tetrachlorobiphenyl |  |
| 73 | 2,3,4,5-Tetrachlorobiphenyl | 3.85 |
| 74 | 2,3', 4, 4',5',6-Hexachlorobiphenyl | 4.00 |

## Geometry optimization and calculation of molecular descriptors

Complete geometry optimization of the 74 molecules of polychlorinated aromatic compounds was performed using Spatan "14"1.1.2 software. Density functional theory (DFT) was used as the level of theory, 6-31G* as the basis set and MMFF as Geometry. The second step in developing the model was the numerical description of molecular structures by defining descriptors. These descriptors were responsible for encoding important features of the structures. A large number of about 1700 molecular descriptors ( $0 \mathrm{D}, 1 \mathrm{D}, 2 \mathrm{D}$ and 3D) were calculated. Quantum chemical descriptors and some of the constitutional descriptors were calculated using Spatan "14"1.1.2 software while topological descriptors and geometrical descriptors were calculated using PaDel-Descriptor 2.18 software.

## Statistical method/correlation analysis

Because of the large number of the descriptors calculated, a stepwise multiple linear regression procedure on the forward-selection and backward-elimination method was used for the selection and elimination of the descriptors. From the square correlation matrix obtained, pairs of variables that falls within the range $0.35 \leq \mathrm{r} \leq 0.9$ were selected and used by the statistical tool to generate the models.

## Development of QSAR models

Genetic Function Algorithm (GFA).
In this work, all the models were developed using genetic function approximation (GFA) technique. The genetic function approximation algorithm was initially anticipated by: (1) Holland's genetic algorithm and (2) Friedman's multivariate adaptive regression splines (MARS) algorithm. In this algorithm, an individual or model is represented as one dimensional string of bits. A distinctive feature of GFA is that it produces a population of several models instead of generating a single model, as do most other statistical methods. Genetic algorithm makes superior models to those developed using stepwise regression techniques because it selects the basis function genetically [23].
The GFA algorithm approach has several important advantages over other techniques: (1) it builds multiple models rather than a single model. (2) It automatically selects which features are to be used in the models. (3) It is better at discovering combinations of features that take advantage of correlations between multiple features. (4) It incorporates Friedman's lack-of fit (LOF) error measure, which estimates the most appropriate number of features, resists over fitting, and allows control over the smoothness of fit. (5) It can use a large variety of equation term types in construction of its models, e.g., splines, step functions, high order polynomials. (6) It provides, through study of
the evolving models, additional information not available from standard regression analysis, such as the preferred model length and useful partitions of the data set [24-25].

QSAR analysis in computational research is responsible for the generation of models to correlate biological activity and physicochemical properties of a series of compounds. The underlying assumption is that the variations of biological activity within a series can be correlated with changes in measured or computed molecular features of the molecules. In the present study, QSAR model generation was performed by GFA technique. The application of the GFA algorithm allows the construction of high-quality predictive models and makes available additional information not provided by standard regression techniques, even for data sets with many features [27-27]. GFA was performed using 100,000 crossovers, smoothness value of 2.0 and other default settings for each combination. The number of terms in the equation was fixed to 10 including constant in the training set. The set of equations generated were evaluated on the basis of some statistical parameters.

## Statistical/Validation Parameter

Lack of fit (LOF)
A "fitness function" or lack of fit (LOF) was used to estimate the quality of the model, so that best model receives the best fitness score. The error measurement term is determine by equation-1
$L O F=\frac{L S E}{\left(1-\frac{c+d * p}{M}\right)^{2}}$
where ' $c$ ' is the number of basic functions (other than constant term); ' $d$ ' is smoothing parameter (adjustable by the user); ' $M$ ' is the number of samples in the training set; LSE is least squares error and ' p ' is the total numbers of the features contained in all basis functions [28].

## Coefficient of multiple determination $\left(R^{2}\right)$

To assess the goodness-of-fit, the coefficient of multiple determination is used. $\mathrm{R}^{2}$ estimates the proportion of the variation in the response that is explained by the predictor.
$R^{2}=1-\frac{\sum_{i=1}^{I}\left(y_{i}-\hat{y}_{i}\right)^{2}}{\sum_{i=1}^{I}\left(y_{i}-\bar{y}\right)}$.
Where yi is the observed dependent variable, $\bar{y}$ the mean value of the dependent variable and $\hat{y}$ the calculated dependent variable. If there is no linear relationship between the dependent variable and the descriptors then $\mathrm{R}^{2}=$ 0.00 ; if there is a perfect fit then $\mathrm{R}^{2}=1.00 . \mathrm{R}^{2}$ values higher than 0.5 indicates that the explained variance by the model is higher the unexplained one.

## Adjusted $R^{2}\left(R^{2}{ }_{\text {adj }}\right)$

The value of $\mathrm{R}^{2}$ can generally be increased by adding additional predictor variables to the model, even if the added variable does not contribute to reduce the unexplained variance of the dependent variable. It follows $\mathrm{R}^{2}$ should be used with caution. This can be avoided by using another statistical parameter the so-called adjusted $R^{2}\left(R^{2}{ }_{\text {ad }}\right)$.
$R_{a d j}^{2}=1-\left(1-R^{2}\right)\left(\frac{I-1}{I-K}\right)$
$R^{2}$ adj is interpreted similarly to the $R^{2}$ value, except that it takes into consideration the number of degrees of freedom. The value of $\mathrm{R}^{2}$ adj decreases if an added variable to the equation does not reduce the unexplained variable.

Standard error of estimate (SEE)
SEE $=\sqrt{\frac{\sum_{i=1}^{I}\left(y_{i}-\hat{y}_{i}\right)^{2}}{(I-(K+1))}}$.
The smaller the value of SEE is, the higher the reliability of the prediction. However, it is not recommended to have the standard error of estimate smaller than the experimental error of the biological data, because it is an indication of over fitted model.

## $F$-value

The F-value is determined using equation-6
$F=\frac{\sum_{i=0}^{I}\left(y_{i}-\bar{y}\right)^{2} /(K-1)}{\sum_{i=1}^{I}\left(y_{i}-\hat{y}_{i}\right)^{2} /(I-K)}$.
The higher the F -value, the greater the probability that the equation is significant [29].

## Validation Parameters

Cross-validation squared correlation coefficient $R^{2}\left(R_{c v}^{2}\right)$
Cross-validation squared correlation coefficient $\mathrm{R}^{2}\left(\mathrm{LOO}-\mathrm{Q}^{2}\right)$ is calculated according to the formula:
$Q^{2}=1-\frac{\sum\left(Y_{\text {pred }}-Y\right)^{2}}{\sum(Y-\bar{Y})^{2}} \ldots$
In Eq. (2), $\mathrm{Y}_{\text {pred }}$ and Y indicate predicted and observed activity values respectively and $\bar{Y}$ indicate mean activity value. A model is considered acceptable when the value of $\mathrm{Q}^{2}$ exceeds 0.5. [30].

In the case of this research, external validation techniques (LMO-Leave Many Out) was applied in which the 23 compounds of the test set were used for the external validation and the predicted $\mathrm{R}^{2}$ for the validation was calculated using equation- 2 .

Predicted $R^{2}\left(R_{\text {pred }}^{2}\right)$
The predictive $\mathrm{R}^{2}$ was calculated based on only molecules not included in the training set (test set). Models are generated based on training set compounds and predictive capacity of the models was judged based on the predictive $\mathrm{R}^{2}\left(\mathrm{R}_{\text {pred }}^{2}\right)$ value which was calculated using eqn -5 .
$R_{\text {pred }}^{2}=1-\frac{\sum\left(Y_{\text {pred }(\text { test })}-Y_{\text {test }}\right)^{2}}{\sum\left(Y_{(\text {test })}-\bar{Y}_{\text {training }}\right)^{2}}$
In Eq. (5), $Y_{\text {pred(test) }}$ and $Y_{\text {test }}$ indicate predicted and observed activity values respectively of the test set compounds and $\bar{Y}_{\text {training }}$ indicates mean activity of the training set. For a QSAR model, the value of $\mathrm{R}^{2}$ pred should be more than 0.5 . All the statistical parameters calculated, agree with the criteria reported in Table-2

Table-2: criteria for selection of good model

| S/N | CRITERIA FOR SELECTION OF MODEL |
| :---: | :---: |
| 1 | $\mathrm{N}=$ number of molecules ( $>20$ molecules) |
| 2 | $\mathrm{K}=$ number of descriptors in a model (statistically $\mathrm{N} / 5$ descriptor in a model) |
| 3 | $\mathrm{df}=$ degree of freedom ( $\mathrm{N}-\mathrm{K}-1$ ) (higher is better). |
| 4 | $\mathrm{R}^{2}=$ coefficient of determination (>0.7) |
| 5 | $\mathrm{R}^{2} \mathrm{cv}=$ cross-validation square correlation (>0.5) |
| 6 | $\mathrm{R}^{2}$ adj = adjusted squared correlation coefficient ( $>0.5$ ) |
| 7 | $\mathrm{R}^{2}$ pred $=$ predicted coefficient of determination ( $>0.5$ ) |
| 8 | SEE $=$ standard error of estimate (smaller is better) |
| 9 | F-test = F-test for statistical significance of the model (higher is better, for some set of descriptors and compounds) |

## RESULTS AND DISCUSSION

## Generation of models

Genetic function approximation was used to performed QSAR regression on 74 molecules of polychlorinated aromatic compounds using $\mathrm{pEC}_{50}$ as dependent variable and calculated molecular descriptors as independent variables described by the equations in Table-2. 51 molecules were used as training set to generate the 5 models which are presented in Table-3.The remaining 23 compounds were used as test set for external prediction and the predicted toxicities are presented in Table-4. Model-1 was selected as the best on the basis of its statistical parameters.

Table- 3.5 generated models by GFA

| No | Equation | Definition |
| :---: | :---: | :---: |
| 1 | $\begin{aligned} & \mathrm{Y}=0.317123855 * \mathrm{X} 26 \\ & +2.673194291 * \mathrm{X} 94 \\ & -1.361102037 * \mathrm{X} 144 \\ & -1.453070371 * \mathrm{X} 165 \\ & -96.631409262 * \mathrm{X} 167 \\ & +69.872692578 * \mathrm{X} 170 \\ & -7.877708429 * \mathrm{X} 228 \\ & -21.439713345 * \mathrm{X} 277 \\ & +124.942300592 * \mathrm{X} 281 \\ & +1.835776438 * \mathrm{X} 293 \\ & -631.436918720 \\ & \hline \end{aligned}$ | X26 : AA : Polarizability X94: CQ : SP-7 <br> X144: EO : maxHBint5 <br> X165 : FJ : ETA_dAlpha_B <br> X167: FL:ETA_Epsilon_2 <br> X170 : FO : ETA_Epsilon_5 <br> X228: HU : n5Ring <br> X277 : JR: GRAV-2 <br> X281: JV : GRAVH-3 <br> X293 : KH: MOMI-R |
| 2 | $\begin{aligned} & \mathrm{Y}=0.337543195 * \mathrm{X} 26 \\ & +1.986019421 * \mathrm{X} 94 \\ & -1.841849479 * \mathrm{X} 144 \\ & -2.591615577 * \mathrm{X} 165 \\ & -105.290996384 * \text { X167 } \\ & +150.560145568 * \mathrm{X} 175 \\ & -8.806840418 * \mathrm{X} 228 \\ & -22.985643463 * \text { X277 } \\ & +135.442609778 * \mathrm{X} 281 \\ & +0.491497214 * \mathrm{X} 45 \\ & -657.102081723 \end{aligned}$ | X26 : AA : Polarizability X94: CQ : SP-7 <br> X144: EO : maxHBint5 <br> X165 : FJ : ETA_dAlpha_B <br> X167 : FL: ETA_Epsilon_2 <br> X175 : FT : ETA_Psi_1 <br> X228: HU : n5Ring <br> X277: JR: GRAV-2 <br> X281: JV: GRAVH-3 <br> X345: MH:WA.eneg |
| 3 | $\begin{aligned} & \mathrm{Y}=0.312654553 * \mathrm{X} 26 \\ & +2.629675114 * \mathrm{X} 94 \\ & -1.401918491 * \mathrm{X} 144 \\ & -1.502764637 * \mathrm{X} 165 \\ & -116.167215005 * \mathrm{X} 167 \\ & +66.429804381 * \mathrm{X} 175 \\ & -7.847220091 * \mathrm{X} 228 \\ & -21.154896743 * \mathrm{X} 277 \\ & +125.332874970 * \mathrm{X} 281 \\ & +1.696580396 * \mathrm{X} 293 \\ & -602.797141503 \end{aligned}$ | X26 : AA : Polarizability X94: CQ : SP-7 <br> X144: EO : maxHBint5 <br> X165 : FJ : ETA_dAlpha_B <br> X167 : FL: ETA_Epsilon_2 <br> X175 : FT : ETA_Psi_1 <br> X228: HU : n5Ring <br> X277: JR: GRAV-2 <br> X281: JV : GRAVH-3 <br> X293 : KH : MOMI-R |
| 4 | $\begin{aligned} & \mathrm{Y}=0.336800494 * \mathrm{X} 26 \\ & +2.942208255 * \mathrm{X} 94 \\ & -0.640140511 * \mathrm{X} 108 \\ & -1.579876967 * \text { X144 } \\ & -1.116608578 * \text { X165 } \\ & +81.907227030 * \mathrm{X} 175 \\ & -8.260828699 * \text { X228 } \\ & -14.048865479 * \text { X277 } \\ & +82.359135734 * \mathrm{X} 281 \\ & +2.468795662 * \text { X293 } \\ & -413.254169614 \end{aligned}$ | X26 : AA : Polarizability X94: CQ : SP-7 <br> X108: DE: nHBa <br> X144: EO : maxHBint5 <br> X165 : FJ : ETA_dAlpha_B <br> X175 : FT : ETA_Psi_1 <br> X228: HU : n5Ring <br> X277 : JR : GRAV-2 <br> X281: JV : GRAVH-3 <br> X293 : KH : MOMI-R |
| 5 | $\begin{aligned} & \mathrm{Y}=0.334445535 * \mathrm{X} 26 \\ & -1.292560192 * \mathrm{X} 35 \\ & +2.861556067 * \text { X94 } \\ & -1.304920784 * \text { X144 } \\ & -1.234903575 * \text { X165 } \\ & -136.309059489 * \text { X177 } \\ & -8.316100485 * \text { X228 } \\ & -17.174253798 * \text { X277 } \\ & +96.839460363 * \text { X281 } \\ & +2.237066146 * \text { X293 } \\ & -454.420645683 \end{aligned}$ | X26 : AA : Polarizability <br> X35 : AJ : apol <br> X94: CQ : SP-7 <br> X144: EO : maxHBint5 <br> X165 : FJ : ETA_dAlpha_B <br> X177 : FV : ETA_dPsi_B <br> X228: HU : n5Ring <br> X277 : JR : GRAV-2 <br> X281: JV : GRAVH-3 <br> X293 : KH : MOMI-R |

Table-4: the best model selected

| Model 1 | Variables in the model |
| :--- | :--- |
| $\mathrm{Y}=0.317123855 *$ X26 | X26 : Polarizability |
| $+2.673194291 *$ X94 | X94 : SP-7 |
| $-1.361102037 *$ X144 | X144 : maxHBint5 |
| $-1.453070371 *$ X165 | X165 : ETA_dAlpha_B |
| $-96.631409262 *$ X167 | X167 : ETA_Epsilon_2 |
| $+69.872692578 *$ X170 | X170 : ETA_Epsilon_5 |
| $-7.877708429 *$ X228 | X228 : n5Ring |
| $-21.439713345 *$ X277 | X277 : GRAV-2 |
| $+124.942300592 *$ X281 | X281: GRAVH-3 |
| $+1.835776438 *$ X293 | X293 : MOMI-R |
| -631.436918720 |  |

Table-5.Chemical names along with the observed and calculated toxicity values in $\mathbf{p E C}_{50}$ of the training set compounds

| S/N | Chemical name | Actual values | Predicted values | Residual values |
| :---: | :---: | :---: | :---: | :---: |
| 1 | 1,2,3,7,8 pentachlorodibenzo-p-dioxin | 7.10000000 | 6.89480200 | 0.205198 |
| 2 | 2,3,6,7 Tetrachlorodibenzo-p-dioxin | 6.80000000 | 6.57743100 | 0.222569 |
| 3 | 2,3,6 Trichlorodibenzo-p-dioxin | 6.66000000 | 6.76089900 | -0.100899 |
| 4 | 1,2,3,4,7,8Hexachlorodibenzo-p-dioxin | 6.55000000 | 6.44193700 | 0.108063 |
| 5 | 1,3,7,8 Tetrachlorodibenzo-p-dioxin | 6.10000000 | 6.11652600 | -0.016526 |
| 6 | 1,2,4,7,8 pentachlorodibenzo--dioxin | 5.96000000 | 6.19664600 | -0.236646 |
| 7 | 1,2,3,4 Tetrachlorodibenzo-p-dioxin | 5.89000000 | 6.28801600 | -0.398016 |
| 8 | 2,3,7 Trichlorodibenzo-p-dioxin | 7.15000000 | 6.86770100 | 0.282299 |
| 9 | 1,2,4 Trichlorodibenzo-p-dioxin | 4.89000000 | 4.88714900 | 0.002851 |
| 10 | 2,8-dichlorodibenzo-p-dioxin | 5.49000000 | 5.15069900 | 0.339301 |
| 11 | 1,2,3,4,6,7,8,9-Octachlorodibenzo-o-dioxin | 5.00000000 | 5.14484700 | -0.144847 |
| 12 | 1-chlorodibenzo-p-dioxin | 4.00000000 | 4.24100500 | -0.241005 |
| 13 | 2,3,7,8 Tetrabromodibenzo-p-dioxin | 8.82000000 | 9.40408400 | -0.584084 |
| 14 | 2,3-Dibromo 7,8-chlorodibenzo-p-dioxin | 8.83000000 | 8.99649900 | -0.166499 |
| 15 | 2,8- Dibromo -3,7-dichlorodibenzo-p-dioxin | 9.35000000 | 9.06707700 | 0.282923 |
| 16 | 2-Bromo-3,7,8-trichlorodibenzo-p-dioxin | 7.94000000 | 8.30822700 | -0.368227 |
| 17 | 1,3,7,9-Tetrabromodibenzo-p-dioxin | 7.03000000 | 7.26343800 | -0.233438 |
| 18 | 1,3,7,8-Tetrabromodibenzo-p-dioxin | 8.70000000 | 8.47020000 | 0.229800 |
| 19 | 1,2,4,7,8-Pentabromodibenzo-p-dioxin | 7.77000000 | 7.42604000 | 0.343960 |
| 20 | 1,2,3,7,8-Pentabromodibenzo-p-dioxin | 8.18000000 | 8.09153100 | 0.088469 |
| 21 | 2,7-Dibromodibenzo-p-dioxin | 7.81000000 | 7.82283900 | -0.012839 |
| 22 | 2-Bromodibenzo-p-dioxin | 6.53000000 | 6.15431200 | 0.375688 |
| 23 | 1-chlorodibenzofuran | 4.53000000 | 4.36064600 | 0.169354 |
| 24 | 2-chlorodibenzofuran | 3.55000000 | 4.22833800 | -0.678338 |
| 25 | 3-chlorodibenzofuran | 4.38000000 | 4.10933400 | 0.270666 |
| 26 | 4-chlorodibenzofuran | 3.00000000 | 3.01669200 | -0.016692 |
| 27 | 2,3-Dichlorodibenzofuran | 5.36000000 | 5.63232200 | -0.272322 |
| 28 | 1,3,6-Trichlorodibenzofuran | 5.36000000 | 5.30147400 | 0.058526 |
| 29 | 1,3,8-Trichlorodibenzofuran | 4.07000000 | 4.56537300 | -0.495373 |
| 30 | 2,3,7-Trichlorodibenzofuran | 7.10000000 | 6.96629000 | 0.133710 |
| 31 | 2,3,4,8-Tetrachlorodibenzofuran | 6.70000000 | 6.40437800 | 0.295622 |
| 32 | 2,3,7,8-Tetrachlorodibenzofuran | 7.39000000 | 7.77433500 | -0.384335 |
| 33 | 1,2,3,7,8-Pentachlorodibenzofuran | 7.13000000 | 6.97671300 | 0.153287 |
| 34 | 1,2,4,7,8-Pentachlorodibenzofuran | 5.89000000 | 5.69544900 | 0.194551 |
| 35 | 2,3,4,7,8-Pentachlorodibenzofuran | 7.82000000 | 7.82693800 | -0.006938 |
| 36 | 1,2,3,4,7,8-Hexachlorobenzofuran | 6.64000000 | 6.83519100 | -0.195191 |
| 37 | 1,2,3,6,7,8-Hexachlorobenzofuran | 6.57000000 | 6.76563400 | -0.195634 |
| 38 | 2,3,4,6,7,8-Hexachlorobenzofuran | 7.33000000 | 7.51370700 | -0.183707 |
| 39 | 2,3,6,8-Tetrachlorodibenzofuran | 6.66000000 | 6.24406500 | 0.415935 |
| 40 | 1,2,3,7-Tetrachlorodibenzofuran | 6.96000000 | 6.86559800 | 0.094402 |
| 41 | 2,3,4,7,9-Pentachlorodibenzofuran | 6.70000000 | 6.37662100 | 0.323379 |
| 42 | 1,2,3,7,9-Pentachlorodibenzofuran | 6.40000000 | 6.39655100 | 0.003449 |
| 43 | Dibenzofuran | 3.00000000 | 2.89483700 | 0.105163 |
| 44 | 2,3,4,7-Tetrachlorobiphenyl | 7.60000000 | 7.48510400 | 0.114896 |
| 45 | 1,2,4,6,8-Pentachlorobiphenyl | 5.51000000 | 5.31940200 | 0.190598 |
| 46 | 2,3,4,4'-Tetrachlorobiphenyl | 4.94000000 | 4.94644900 | -0.006449 |
| 47 | 3,3',4,4'-Tetrachlorobiphenyl | 6.15000000 | 6.06549900 | 0.084501 |
| 48 | 3,3'4,4'5-Pentachlorobiphenyl | 6.92000000 | 6.42820400 | 0.491796 |
| 49 | 2,3,3',4,4',5-Hexachlorobiphenyl | 5.15000000 | 5.63058300 | -0.480583 |
| 50 | 2,3,3',4,4',5'-Hexachlorobiphenyl | 5.30000000 | 5.42050100 | -0.120501 |
| 51 | 2,3,4,5-Tetrachlorobiphenyl | 3.85000000 | 3.89186900 | -0.041869 |

Table-6. Chemical names along with the observed and the calculated toxicities of the test set compounds expressed in $\log ^{\mathbf{5 E C}} \mathrm{C}_{50}$

| S/N | Compound names | Actual values | Predicted values | Residual values |
| :---: | :---: | :---: | :---: | :---: |
|  |  |  |  |  |
| 1 | 2,3,7,8 Tetrachlorodibenzo-p-dioxin | 8.00 | 7.137742 | 0.862258 |
| 2 | 1,2,3,4,7 pentachlorodibenzo-p-dioxin | 5.19 | 6.416795 | -1.2468 |
| 3 | 2,3,7-Tribromodibenzo-p-dioxin | 8.93 | 9.433407 | -0.50341 |
| 4 | 2,6-dichlorodibenzofuran | 3.61 | 5.33345 | -1.72345 |
| 5 | 2,8-Dichlorodibenzofuran | 5.05 | 5.814313 | -0.76431 |
| 6 | 2,3,4-Trichlorodibenzofuran | 4.72 | 6.341931 | -1.62193 |
| 7 | 2,3,8-Trichlorodibenzofuran | 6.00 | 6.078967 | -0.07897 |
| 8 | 2,6,7 Trichlorodibenzofuran | 6.35 | 5.891172 | 0.458828 |
| 9 | 2,3,4,6 Tetrachlorodibenzofuran | 6.46 | 6.643481 | -0.18348 |
| 10 | 1,2,4,8-Tetrachlorodibenzofuran | 5.00 | 5.764051 | -0.76405 |
| 11 | 1,2,4,7,9-Pentachlorodibenzofuran | 4.7 | 4.070704 | 0.629296 |
| 12 | 1,2,4,6,7,9-Hexachlorodibenzofaran | 5.08 | 3.868581 | 1.211419 |
| 13 | 1,2,3,6-Tetrachlorodibenzofuran | 6.46 | 6.216 | 0.244 |
| 14 | 1,3,4,7,8-Pentachlorodibenzofuran | 6.70 | 6.636922 | 0.063078 |
| 15 | 3,4,4',5-Tetrachlorobiphenyl | 4.55 | 5.973791 | -1.42379 |
| 16 | $2^{\prime}, 3,4,4$ ',5-Pentachlorobiphenyl | 4.85 | 4.9348 | -0.0848 |
| 17 | 2,3,3',4,4'-Pentachlorobiphenyl | 5.37 | 5.668328 | -0.29833 |
| 18 | 2,3',4,4',5-Pentachlorobiphenyl | 5.04 | 5.392424 | -0.35242 |
| 19 | 2,3,4,4',5-Pentachlorobiphenyl | 5.39 | 5.43461 | -0.04461 |
| 20 | 2,2'4,4',5,5'-Hexachlorobiphenyl | 4.26 | 4.397758 | -0.13776 |
| 21 | 2,3',4,4',5,5'-Hexachlorobiphenyl | 4.80 | 4.397758 | 0.402242 |
| 22 | 2,2',4,4'-Tetrachlorobiphenyl | 3.89 | 4.710143 | -0.82014 |
| 23 | 2,3', 4, $4^{\prime}, 5^{\prime}, 6$-Hexachlorobiphenyl | 4.00 | 3.366028 | 0.633972 |

Table-7. Statistical/validation parameters of the generated models

| Statistical parameters | Model 1 | Model 2 | Model 3 | Model 4 | Model 5 |
| :--- | :--- | :--- | :--- | :--- | :--- |
| Friedman LOF | 0.437664 | 0.439352 | 0.440601 | 0.443745 | 0.443884 |
| R-squared | 0.967342 | 0.967217 | 0.967123 | 0.966889 | 0.966878 |
| Adjusted R-squared | 0.959178 | 0.959021 | 0.958904 | 0.958611 | 0.958598 |
| Cross validated R-squared | 0.940154 | 0.923084 | 0.937807 | 0.933100 | 0.937382 |
| Significant Regression | Yes | Yes | Yes | Yes | Yes |
| Significance-of-regression F-value | 118.48328 | 118.012638 | 117.666633 | 116.80482 | 116.7668 |
| Critical SOR F-value (95\%) | 2.080618 | 2.080618 | 2.080618 | 2.080618 | 2.080618 |
| Replicate points | 0 | 0 | 0 | 0 | 0 |
| Computed experimental error | 0.000000 | 0.000000 | 0.000000 | 0.000000 | 0.000000 |
| Lack-of-fit points | 40 | 40 | 40 | 40 | 40 |
| Min expt. error for non-significant LOF $(95 \%)$ | 0.257926 | 0.258423 | 0.258790 | 0.259712 | 0.259752 |

Table-8: Univariate analysis of the toxicity data

| Statistical parameter | values |
| :--- | :--- |
| Number of sample points | 51 |
| Range | 6.35000000 |
| Maximum | 9.35000000 |
| Minimum | 3 |
| Mean | 6.28450980 |
| Median | 6.57000000 |
| Variance | 2.23024000 |
| Standard deviation | 1.50826000 |
| Mean absolute deviation | 1.21134000 |
| Skewness | -0.26949300 |
| Kurtosis | -0.48857800 |

Table-9 values of the descriptors used in the selected model

| S/N | p | SP-7 | minHBint5 | ETA_dAlpha_B | ETA_Epsilon_2 | ETA_Epsilon_5 | n5Ring | GRAV-2 | GRAVH-3 | MOMI-R |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
| 1 | 61.63 | 7.185 | 0.8984 | 22.9790 | 0.5291 | 0.871 | 0.4444 | 42.242 | 12.218 | 6.9852 |
| 2 | 61.63 | 7.849 | 0.9207 | 27.0342 | 0.53885 | 0.871 | 0.4444 | 43.892 | 12.507 | 7.2141 |
| 3 | 59.39 | 8.003 | 0.9321 | 27.321 | 0.53885 | 0.7723 | 0.4444 | 43.893 | 12.506 | 7.0553 |
| 4 | 62.75 | 6.738 | 0.8647 | 19.581 | 0.51821 | 0.9203 | 0.375 | 40.525 | 11.916 | 6.4812 |
| 5 | 60.51 | 8.526 | 0.9543 | 31.143 | 0.54762 | 0.8216 | 0.4444 | 45.483 | 12.782 | 7.4407 |
| 6 | 61.63 | 6.983 | 0.9049 | 23.201 | 0.5291 | 0.871 | 0.4444 | 42.243 | 12.218 | 6.9739 |
| 7 | 59.4 | 7.734 | 0.9322 | 27.2260 | 0.53885 | 0.7722 | 0.4444 | 43.893 | 12.507 | 7.2520 |
| 8 | 60.53 | 6.600 | 0.8414 | 19.2939 | 0.51821 | 0.8216 | 0.4444 | 40.525 | 11.916 | 6.4875 |
| 9 | 59.38 | 7.604 | 0.8636 | 23.8245 | 0.5291 | 0.7723 | 0.375 | 42.242 | 12.218 | 6.4679 |
| 10 | 58.28 | 7.940 | 0.8973 | 27.4095 | 0.53885 | 0.7229 | 0.4444 | 43.893 | 12.507 | 7.0710 |
| 11 | 64.96 | 6.813 | 0.8413 | 19.9452 | 0.51821 | 1.0191 | 0.5 | 40.525 | 11.915 | 6.2940 |
| 12 | 57.13 | 6.015 | 0.8191 | 15.6297 | 0.50595 | 0.6736 | 0.5 | 38.732 | 11.597 | 6.2049 |
| 13 | 61.99 | 9.866 | 1.1305 | 39.4347 | 0.5628 | 0.7091 | 0.4444 | 48.507 | 13.301 | 7.6460 |
| 14 | 61.25 | 5.817 | 0.7918 | 12.3852 | 0.49206 | 0.7654 | 0.4286 | 36.851 | 11.259 | 5.4654 |
| 15 | 61.24 | 7.185 | 0.9407 | 14.8188 | 0.66667 | 0.7654 | 0.4444 | 47.829 | 13.252 | 7.9693 |
| 16 | 60.88 | 7.185 | 0.9119 | 18.8989 | 0.59788 | 0.7935 | 0.4444 | 45.122 | 12.756 | 7.4719 |
| 17 | 61.98 | 7.185 | 0.9154 | 18.899 | 0.59788 | 0.7091 | 0.4444 | 45.122 | 12.756 | 7.5033 |
| 18 | 61.82 | 7.185 | 0.9042 | 20.9390 | 0.56349 | 0.7091 | 0.4444 | 43.706 | 12.493 | 7.2480 |
| 19 | 63.51 | 6.781 | 0.9790 | 15.0493 | 0.66667 | 0.7303 | 0.5 | 47.829 | 13.251 | 7.9030 |
| 20 | 63.55 | 6.983 | 0.9493 | 14.9323 | 0.66667 | 0.7303 | 0.4444 | 47.829 | 13.252 | 7.9577 |
| 21 | 59.01 | 7.734 | 0.9857 | 16.8090 | 0.70175 | 0.6667 | 0.4444 | 50.553 | 13.725 | 8.2323 |
| 22 | 57.5 | 7.850 | 0.9708 | 16.7255 | 0.70175 | 0.6455 | 0.4444 | 50.552 | 13.724 | 8.242 |
| 23 | 56.4 | 6.600 | 0.8661 | 13.1738 | 0.62745 | 0.6358 | 0.4444 | 44.941 | 12.742 | 7.4339 |
| 24 | 56.42 | 6.015 | 0.8446 | 11.5452 | 0.58333 | 0.6358 | 0.4444 | 41.854 | 12.189 | 6.4238 |
| 25 | 56.42 | 5.818 | 0.7997 | 10.2368 | 0.53333 | 0.6358 | 0.4285 | 38.520 | 11.580 | 5.6197 |
| 26 | 50.4 | 5.504 | 0.9954 | 8.35188 | 0.5034 | 0.6358 | 0.3333 | 36.569 | 11.204 | 5.0549 |
| 27 | 57.55 | 5.433 | 0.9840 | 8.02229 | 0.5034 | 0.6875 | 0.4285 | 36.569 | 11.204 | 5.2679 |
| 28 | 58.66 | 5.427 | 0.9840 | 7.96089 | 0.5034 | 0.7392 | 0.4285 | 36.569 | 11.204 | 5.1256 |
| 29 | 58.7 | 5.570 | 0.9954 | 8.26993 | 0.5034 | 0.7392 | 0.3333 | 36.569 | 11.204 | 5.1663 |
| 30 | 58,7 | 6.019 | 1.0063 | 11.6458 | 0.51746 | 0.7392 | 0.4286 | 38.464 | 11.545 | 5.6225 |
| 31 | 59.81 | 5.913 | 1.0291 | 11.8161 | 0.51746 | 0.7909 | 0.4285 | 38.463 | 11.545 | 5.7123 |
| 32 | 59.82 | 5.776 | 1.0177 | 11.6102 | 0.51746 | 0.7909 | 0.4286 | 38.464 | 11.545 | 6.0089 |
| 33 | 60.93 | 6.228 | 1.1096 | 15.8458 | 0.52976 | 0.8426 | 0.4286 | 40.269 | 11.867 | 6.1781 |
| 34 | 60.94 | 6.091 | 1.0514 | 15.6334 | 0.52976 | 0.8426 | 0.5 | 40.269 | 11.867 | 6.1080 |
| 35 | 60.78 | 6.689 | 1.0399 | 15.7116 | 0.52976 | 0.8426 | 0.4285 | 40.269 | 11.867 | 5.9662 |
| 36 | 62.05 | 6.354 | 1.0634 | 15.2198 | 0.52976 | 0.8943 | 0.375 | 40.269 | 11.866 | 6.1495 |
| 37 | 62.04 | 6.361 | 1.0399 | 15.3223 | 0.52976 | 0.8943 | 0.5 | 40.269 | 11.866 | 6.2819 |
| 38 | 62.04 | 6.513 | 1.0513 | 15.5286 | 0.52976 | 0.8943 | 0.5 | 40.269 | 11.866 | 6.0816 |
| 39 | 59.81 | 7.168 | 1.1589 | 19.6480 | 0.54062 | 0.7909 | 0.4286 | 41.997 | 12.171 | 6.3935 |
| 40 | 59.81 | 7.031 | 1.0736 | 19.4259 | 0.54062 | 0.7909 | 0.5 | 41.996 | 12.172 | 6.5899 |
| 41 | 60.93 | 6.946 | 1.0970 | 19.0551 | 0.54062 | 0.8426 | 0.375 | 41.996 | 12.171 | 6.6103 |
| 42 | 60.93 | 6.853 | 1.0850 | 19.8096 | 0.54062 | 0.8426 | 0.4285 | 41.996 | 12.172 | 6.4965 |
| 43 | 55.27 | 7.174 | 1.1656 | 23.9134 | 0.55027 | 0.5841 | 0.5 | 43.656 | 12.462 | 6.6331 |
| 44 | 59.81 | 7.548 | 1.1307 | 23.3562 | 0.55027 | 0.7909 | 0.375 | 43.656 | 12.462 | 6.6939 |
| 45 | 60.95 | 7.438 | 1.1421 | 23.5761 | 0.55027 | 0.8426 | 0.5 | 43.656 | 12.462 | 6.8714 |
| 46 | 59.83 | 7.616 | 1.1307 | 23.1860 | 0.55027 | 0.8426 | 0.375 | 43.656 | 12.462 | 6.9490 |
| 47 | 59.97 | 8.230 | 1.1644 | 27.5404 | 0.5589 | 0.7156 | 0.375 | 45.255 | 12.740 | 7.0311 |
| 48 | 61.09 | 8.218 | 1.1878 | 27.5241 | 0.5589 | 0.7649 | 0.375 | 45.255 | 12.740 | 7.0098 |
| 49 | 62.16 | 8.287 | 1.2497 | 27.3540 | 0.5589 | 0.8143 | 0.375 | 45.255 | 12.740 | 7.1719 |
| 50 | 62.04 | 7.930 | 1.2611 | 28.0611 | 0.5589 | 0.8143 | 0.5 | 45.255 | 12.740 | 6.8331 |
| 51 | 59.83 | 6.750 | 1.1085 | 19.3050 | 0.54062 | 0.7155 | 0.5 | 41.996 | 12.172 | 6.8053 |

Table-10: definition of descriptors that were found in the 5 models

| Descriptor | Definition |
| :--- | :--- |
| Polarizability | Determine the dynamical response of a bound system to external fields, and provide insight into a molecule's internal <br> structure. |
| SP-7 | Chi path descriptor with a simple path order 7 |
| MaxHBint5 | Maximum E-state descriptor of strength for potential hydrogen bonds of path length 5 |
| ETA_dApha_B | Extended Topochemical Atomic descriptor which is defined as a measure of count of hydrogen bond acceptor atoms and/or <br> polar surface area |
| ETA_Epsilon_2 | Extended Topochemical Atomic descriptor which is defined as a measure of electronegative atom count 2 |
| ETA_Epsilon_5 | Extended Topochemical Atomic descriptor which is defined as a measure of electronegative atom count 5 |
| n5Ring | Ring count descriptor which indicates 5 member rings |
| GRAV_2 | Gravitational index descriptor which is defined as square root of gravitational index of heavy atom |
| GRAVH_3 | Gravitational index descriptor which is defined as cube root of hydrogen-included gravitational index |
| MOMI-R | Moment of initia along the radius of gyration |
| ETA_psi_1 | Measure of hydrogen bond propensity the molecules and/or polar surface area. |
| WA.eng | Non directional WHIM, weighted by Mulliken atomic electronegativites |
| apol | Sum of atomic polarizabilities (including implicit hydrogen) |
| nHBa | Electrotopological state atom type descriptor which is defined as count of E-state for hydrogen bond acceptors |
| ETA_dpsi_B | Measure of hydrogen bonding propensity of the molecules |

Table-11: The definition of the descriptors used in model-1 and their regression coefficients

| Descriptor <br> notation | Definition | Regression <br> coefficient |
| :--- | :--- | :--- |
| Polarizability (p) | Determine the dynamical response of a bound system to external fields, and provide insight into a <br> molecule's internal structure. | 0.317123855 |
| SP-7 | Chi path descriptor with a simple path order 7 | 2.673194291 |
| MaxHBint5 | Maximum E-state descriptor of strength for potential hydrogen bonds of path length 5 | -1.36110204 |
| ETA_dApha_B | Extended Topochemical Atomic descriptor which is defined as a measure of count of hydrogen bond <br> acceptor atoms and/or polar surface area | -1.45307037 |
| ETA_Epsilon_2 | Extended Topochemical Atomic descriptor which is defined as a measure of electronegative atom <br> count 2 | -96.6314093 |
| ETA_Epsilon_5 | Extended Topochemical Atomic descriptor which is defined as a measure of electronegative atom <br> count 5 | 69.8726926 |
| n5Ring | Ring count descriptor which indicates 5 member rings | -7.87770843 |
| GRAV_2 | Gravitational index descriptor which is defined as square root of gravitational index of heavy atom | -21.4397133 |
| GRAVH_3 | Gravitational index descriptor which is defined as cube root of hydrogen-included gravitational index | 124.9423006 |
| MOMI-R | Moment of initia along the radius of gyration | 1.835776438 |



Fig. 2: Linear relationship of observed and predicted toxicities of data of the training set


Fig. 3: Linear relationship of observed and predicted toxicities of data of the test set


Fig.4. residual versus actual values

## DISCUSSION

Table-3 shows 5 models generated by GFA using all the compounds of the training set. The toxicity (Y) was used as independent variables and the descriptors ( Xi ) as dependent variables. Each model contains 10 descriptors as this agrees with the second criteria reported in Table-2. Among the 5 models generated, model-1 in Table-4 was selected as the best on the basis of the various statistical parameters reported in Table-7. Table-5 shows the predicted toxicities of the training set in pEC50 which has a good agreement with the experimental toxicities. Table-6 reported the predicted toxicities of test set for external validation with which the predicted ( $\mathrm{R}_{\text {pred. }}^{2}$ ) was calculated as 0.7209 and is in excellent agreement with criteria reported in Table-2

Table-7 shows the statistical/validation parameters of all the 5 models. The statistical quality of the models were determined by the validation parameters like LOF which is the measure of quality of fit, $\mathrm{R}^{2}, \mathrm{R}_{\mathrm{adj}}{ }^{2}, \mathrm{R}^{2}{ }_{\mathrm{cv}}, \mathrm{F}$-test and the larger the value the better the model and the external validation parameter $\mathrm{R}_{\text {pred. }}^{2}$. The statistical parameters
model-1 appear to be statistically better than those of the other models. For model-1: $\mathrm{LOF}=0.4377, \mathrm{R}_{2}=0.9673$, $\mathrm{R}^{2}{ }_{\text {adj. }}=0.9592, \mathrm{R}_{2 \mathrm{cv}}=0.9402$, F-test $=118.48$ and $\mathrm{R}_{\text {pred. }}^{2}=0.7209$. All of these parameters are in a very good agreement with criteria reported in Table-2

Table-8 shows the statistical parameters of univariate analysis that describe the toxicity data. The most important parameters here are skewness and kurtosis. Skewness is the third moment of the distribution, which indicates the symmetry of distribution. As skewness is positive, the distribution of data value within the column is skewed toward positive values. For a symmetry distribution, the skewness is close to zero. Kurtosis is the fourth moment of the distribution which indicates the profile of the column of data relative to normal distribution [32].

## Descriptor contribution

(Huifeng et al., 2011) reported that radius of gyration (RGyr) and Ist component accessibilities directional index/weight by atomic polarizabilities ( $\mathrm{E}_{\mathrm{Ip}}$ ) are among the descriptors that are responsible for producing toxicties of polychlorinated aromatic compounds. Another previous work by (Nandan et al., 2013) shown that the descriptors Winner index (W), Balban index (J), polarizability ( $\alpha$ ) and index of refraction ( $\eta$ ) have high responsibilities in producing toxicity of some polychlorinated aromatic compounds.

The present QSAR model study reveals that apart from the descriptors reported by (Huifeng et al., 2011; Nandan et al., 2013) which are responsible for producing toxicity of polychlorinated aromatic compounds, other descriptors were also found to be responsible for producing toxicity of polychlorinated aromatic compounds. Among these descriptors, Polarizability, SP-7, ETA_Epsilon_5, GRAVH_3, and MOMI-R which are used in model-1 contribute positively in producing toxicities of polychlorinated aromatic compounds. This indicate a positive impact on the toxicities of polychlorinated aromatic compounds, which means increasing the value of this descriptors produces higher toxicities of these compounds. In the other hand, the descriptors maxHBint5, ETA_dApha_B, ETA_Epsilon_2, n5Ring and GRAV_2 with negative coefficient used in model-1 contribute negatively, hence decreasing the values of these descriptors will provide higher toxicities of polychlorinated aromatic compounds. The interpretation of this model shows that each of these descriptors with positive coefficient is directly proportional to the toxicities of these molecules while each of those descriptors with negative coefficient is inversely proportional to the toxicities of the molecules [33]. Model-1 is presented in Table-4 and the descriptors used in model-1 are listed in Table-8. It is observed that both in this work and the once reported by (Huifeng et al., 2011; Nandan et al., 2013 polarizability and radius of gyration ( RGyr ) contribute in producing toxicity of polychlorinated aromatic compounds.

Figure-2 shows a plot describing the linear relationship between the experimental values in $\mathrm{pEC}_{50}$ and the calculated values. Most of the compounds of the training set are along the linear line of the plot. This indicates that the predicted values of $\mathrm{pEC}_{50}$ are in agreement with the experimental values. But for the test set, whose imprecise toxicity data were reported as shown in Figure-3, errors are higher than the training set. Figure 4 shows the plot of residuals versus experimental values of data set. The propagation of residuals on both sides of zero indicates that no systematic error exists in the development of GFA.

## CONCLUSION

A genetic function approximation method was used to run the regression analysis and establish correlation's between different types of descriptors and experimental toxicity of three classes of polychlorinated aromatic (PCDDs, PCDFs and PCBs). QSAR models were developed and one of them was used to predict the toxicity efficiency of polychlorinated aromatic compounds. The prediction of toxicity efficiencies of these compounds matched with the experimental measurements. The developed models were found to be statistically significant as evidenced from their regression statistics.

Out of about 1700 molecular descriptors generated only these few were found to be the once responsible for producing toxicity of poly. These descriptors include: polalizability, Chi path descriptor with a simple path length order 7 (SP-7), Extended Topochemical Atomic descriptor (ETA_Epsilon_5) which is the measure of electronegative atom count 5, cubic root of hydrogen-included gravitational index (GRAVH-3) and moment of initia along the radius of gyration (MOMI-R). All the calculated molecular descriptors were aimed to encode some important information about the structural features of polychlorinated aromatic compounds which could influence the receptor binding affinity. Some of them provided good correlations and statistically reliable models.

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